

Synthesis of Functionalized Phenolformaldehyde Polymer Resins by the Reaction of 2,4-Dihydroxyacetophenone Formaldehyde Resin with Various Amines and Their Metal Ion Uptake Properties

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Received 23 May 2003; accepted 14 October 2003

ABSTRACT: A series of functionalized phenolformaldehyde polymer resins have been synthesized by the reaction of 2,4-dihydroxyacetophenone-formaldehyde resin with the amines, such as ethanolamine, aminophenol, ethylenediamine, and propylenediamine in dichloromethane. The Schiffbase polymers were characterized by IR and $^1\text{H-NMR}$ spectroscopic techniques. Thermal stabilities of the polymers were determined by TG and DTA studies. Heavy and toxic metal ions viz., Pb(II), Hg(II), Cd(II), and Cr(VI) have been removed using these polymer resins. Metal uptake efficiency, reusability, effect of pH, effect of time, and effect

of initial concentration on the metal uptake were also studied. Amount of metal removed by the resin was determined using atomic absorption spectrophotometry. The retention properties were also tested under competitive conditions and were found to be depend strongly on the pH. Elution of metal ions were investigated in acid media. The uptake behaviour of the resins was approximately described by Freundlich's equation. © 2004 Wiley Periodicals, Inc. *J Appl Polym Sci* 92: 1501–1509, 2004

Key words: synthesis; selectivity

INTRODUCTION

The continuous increase of world needs for most of the known metals, the decrease in grade of available ores, and strict environmental regulations, make it interesting to find effective and efficient methods for processing waste solutions containing metal ions, even at very low concentrations. Adsorption techniques employing activated carbon,¹ starch xanthate,² fly ash,³ and surpentine minerals⁴ have been used for the treatment of industrial waste water. Alkaline precipitation⁵ and ion exchange⁶ methods have also been used for the removal of toxic metals. Although many of these methods have been developed and successfully used, their application can cause problems. Some problems can be connected with heterogeneous reactions and interphase transfer. Other problems can arise if aqueous solutions are preferred for the subsequent procedures rather than organic solvents or solid concentrates. In such cases additional procedures are needed, for example, back extraction, desorption, dissolution of solid concentrates, etc., which complicate the analysis and can result in contamination of the sample from the reagent added. Insoluble polymeric

supports are widely investigated and applied for metal recovery from dilute solutions,^{7–15} in which various chelating groups have been incorporated and are attached to the polymer matrix. Syntheses of such functionalized polymers have been accomplished using either polymerization or simple functionalization principles. The former involves the polymerization of monomers containing the desired ligands. The main advantages of these polymeric resins are high chemical and mechanical stability. In continuation^{16–19} of our research, herein we report the synthesis of a set of selective phenolformaldehyde functionalized resins and the removal of toxic/heavy metal ions. In view of analytical potentialities of 2,4-Dihydroxyacetophenone derivatives,^{20,21} it is considered worthwhile to synthesize a functionalized polymer containing this moiety.

EXPERIMENTAL

Materials and methods

2,4-Dihydroxyacetophenone was prepared according to the literature procedure.²² It was recrystallized from distilled water. Thirty-seven percent formalin (Merck) and oxalic acid (Merck) were used without further purification. Ethanolamine, 2-aminophenol, ethylenediamine, and propylenediamine (Merck) were purified before use. Stock solutions of Pb(II), Hg(II), Cd(II), and Cr(VI) were prepared from AR grade lead

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Contract grant sponsor: CSIR, New Delhi; contract grant number: 01(1558)/98/EMR-II.

nitrate, mercuric chloride, cadmium nitrate, and potassium dichromate salts.

Physical measurements

Elemental analyses were performed microanalytically by CDRI, Lucknow. IR spectra were recorded with KBr pellets, using a Beckmann IR-20 spectrophotometer. NMR spectral analyses were carried out in dimethylsulfoxide- d_6 using a Bruker DSX 500 MHz instrument. NETZSCH, Geratebau, GmbH thermal analysing system (TAS) was used to evaluate the thermal stability, decomposition temperature, and kinetic parameters at heating rate of $10^\circ\text{C min}^{-1}$. Metal uptake by the resin was determined using a UNICAM UK Model 839 Atomic Absorption Spectrophotometer [for Pb(II), Cd(II)] and [Cr(VI)] and a Shimadzu UV-160 A spectrophotometer [for Hg(II)].

Synthesis of 2,4-dihydroxyacetophenone formaldehyde resin (DAPF)

A mixture (1 : 1.1) 2,4-dihydroxyacetophenone and formalin solution, 3% (w/w) oxalic acid were taken in a reaction tube, sealed, and placed in an oil bath at 100°C for 24 h. The tube was then cooled to room temperature, desealed, and water was removed by decantation. The solid remaining in the tube was dissolved in dimethylformamide and the solution was added dropwise to a large excess of 10% aqueous sodium chloride solution, with constant stirring. The resin that separated out was collected by filtration, washed several times with distilled water, until free from chloride ions. The resin was dried at 60°C for about 1 h.

Synthesis of chelating resins

2,4-Dihydroxyacetophenone formaldehyde resin (0.0184 mol, 3 g) is condensed with ethanolamine (0.0184 mol, 1.3 mL), aminophenol (0.0184 mol, 2 g), ethylenediamine (0.0092 mol, 0.616 mL), and propylenediamine (0.0092 mol, 0.75 mL) in dichloromethane (20 mL) by stirring on a magnetic stirrer for about 2 h to get 2,4-dihydroxyacetophenone formaldehyde resin-ethanolamine (DAPF-ea), 2,4-dihydroxyacetophenone formaldehyde resin-aminophenol (DAPF-ap), 2,4-dihydroxyacetophenone formaldehyde resin-ethylenediamine (DAPF-en), and 2,4-dihydroxyacetophenone formaldehyde resin-propylenediamine (DAPF-pn), respectively.

Recommended procedure for the removal of metal ions

The effect of pH on metal ion uptake was studied by stirring 2.5 mL of 1×10^{-2} M metal [Pb(II), Hg(II),

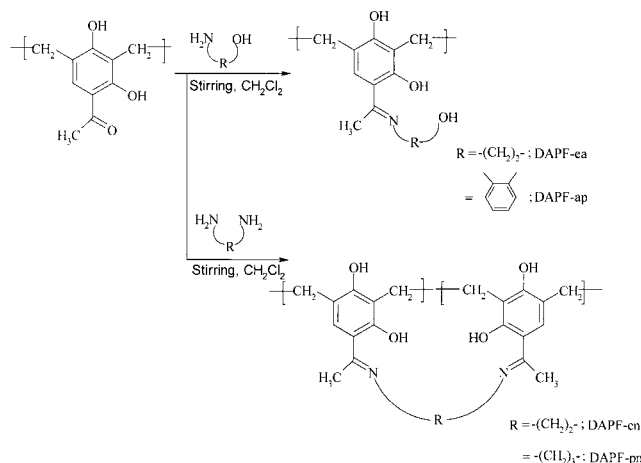


Figure 1 Scheme for the synthesis of chelating polymer resins.

Cd(II) or Cr(VI)] solution, 7.5 mL buffer (pH 2–10) solution and resin 0.2 g in a 100-mL beaker for 1 h. The effect of time was studied by repeating the above procedure at pH 10 at different time intervals (15, 30, 45, 60, and 120 min). The selectivity of the metal ion was determined by stirring an aliquot containing 0.1 M Pb(II) (0.5 mL), Hg(II) (0.5 mL), Cd(II) (1 mL), and Cr(VI) (2 mL) in a 100-mL beaker along with 1 g of the resin and 7.5 mL of buffer (pH 10) solution. The maximum uptake capacities of each polymer resin was determined by reacting 50, 100, 150, 200, 250, and 300 ppm of metal solutions with, polymeric resin under suitable pH conditions [pH 10 for Pb(II) and Hg(II) and pH 6 for Cd(II) and Cr(VI)] for about 1 h. In all cases the metallated resin was filtered and washed thoroughly with distilled water. The filtrate was collected quantitatively. The amount of metal ion present in the filtrate was estimated using atomic absorption spectroscopy and spectrophotometric methods using 1,10-phenanthroline.²³ The amount of metal ion was deduced from the predetermined calibration curve.

RESULTS AND DISCUSSION

The preparative scheme for functionalised polymer resins containing NO_2 and N_2O_2 donor atoms is given in Figure 1. The reaction was carried out in dichloromethane. The 2,4-Dihydroxy phenolformaldehyde resin is colorless while the functionalized resins have yellow and pale yellow colors. The color of these newly formed resins remains unchanged even after several washings with CH_2Cl_2 . Analytical data and yields of the resins are given. DAPF: 92% yield. Anal. calcd: C, 65.85%; H, 4.88%. Found: C, 65.41%, H, 4.9%. DAPF-ea: 83% yield. Anal. calcd: C, 63.76%; H, 6.28%. Found: C, 63.3%; H, 6.3%. DAPF-ap: 85% yield. Anal. calcd: C, 70.58%; H, 5.09%. Found: C, 70.2%; H, 5.1%. DAPF-en: 75% yield. Anal. calcd: C, 68.18%; H, 5.68%.

TABLE I
¹H-NMR Spectral Data of the Resins

Resin	δ (ppm)	Multiplicity	Number of protons	Assignment
DAPF-ea	1.98	Singlet	2H	(—CH ₂ —) of the resin
	2.15	Singlet	3H	(—CH ₃) methyl
	3.3	Triplet	2H	N—CH ₂ — of ethanolamine
	3.7	Triplet	2H	O—CH ₂ — of ethanolamine
	7.45	Singlet	1H	Aromatic
	12.55	Singlet	1H	Phenolic (—OH), <i>ortho</i> to —CH ₂ — group of the resin
	13.05	Singlet	1H	Phenolic (—OH), <i>ortho</i> to condensed ethanolamine
DAPF-ap	1.98	Singlet	2H	(—CH ₂ —) of the resin
	2.15	Singlet	3H	(—CH ₃) methyl
	6.1 to 7.8	Multiplet	5H	Aromatic
	12.3	Singlet	1H	Phenolic (—OH), <i>ortho</i> to —CH ₂ — group of the resin
DAPF-en	13.2	Singlet	1H	Phenolic (—OH), <i>ortho</i> to condensed aminophenol
	1.9	Singlet	2H	(—CH ₂ —) of the resin
DAPF-pn	2.15	Singlet	3H	(—CH ₃) methyl
	3.7	Triplet	2H	(—CH ₂ —) of the condensed ethylene diamine
	7.4	Singlet	1H	Aromatic
	13.1	Singlet	1H	Phenolic (—OH), <i>ortho</i> to —CH ₂ — group of the resin
	13.3	Singlet	1H	Phenolic (—OH), <i>ortho</i> to condensed diamine
	1.95	Singlet	2H	(—CH ₂ —) of the resin
	2.15	Singlet	3H	(—CH ₃) methyl
	2.7	Quintet	2H	(—CH ₂) of propylene diamine
DAPF-pn	3.6	Triplet	2H	(—CH ₂ —) of propylene diamine
	7.5	Singlet	1H	Aromatic
	13.1	Singlet	1H	Phenolic (—OH), <i>ortho</i> to —CH ₂ — group of the resin
	13.3	Singlet	1H	Phenolic (—OH), <i>ortho</i> to condensed propylene diamine

Found: C, 67.3%; H, 5.7%. DAPF-pn: 73% yield. Anal. calcd: C, 68.85%; H, 6.01%. Found: C, 67.8%; H, 6.1%. The data suggest that yields of the resins are sufficiently high. The yields of DAPF-ea and DAPF-ap are found to be better than DAPF-en and DAPF-pn. This is because of stereochemical factors. The analytical data are in good agreement with the calculated values.

The resins are insoluble in common organic solvents and soluble in dimethylformamide and dimethylsulfoxide. ¹H-NMR spectra of the resins were recorded in DMSO-d₆. The multiplicity pattern of the resins are given in Table I. The spectral data support the synthesis of resins. Infrared spectra of the resins were recorded in the range 4000–400 cm⁻¹ using KBr discs. A comparison of IR spectra of the DAPF with the other four functionalized resins, shows the absence of νC=O (1665 cm⁻¹) and the presence of νC=N (azomethine) 1608, 1621, 1614, and 1608 cm⁻¹, respectively, for DAPF-ea, DAPF-ap, DAPF-en, and DAPF-pn. IR; DAPF: ν 3400(O—H), 2885(C—H), 1665 (C=O) cm⁻¹. DAPF-ea: ν 3395(O—H), 2887 (C—H), 1608 (C=N) cm⁻¹. DAPF-ap: ν 3320 (O—H), 2726 (C—H), 1621 (C=N) cm⁻¹. DAPF-en: ν 3380 (O—H), 2609 (C—H), 1614 (C=N) cm⁻¹. DAPF-pn: ν 3394 (O—H), 2856 (C—H), 1608 (C=N) cm⁻¹. The lower OH stretching frequencies of all the resins (3400–3320 cm⁻¹) are due to hydrogen bonding in polymeric association. Thus the above data support the formation of functionalized polymers.

Thermogravimetric analyses

Thermograms of the functionalized polymers are shown in Figure 2. A mathematical interpretation of thermogravimetric curves enables one to determine kinetic parameters of pyrolysis reactions. Horowitz and Metzger²⁴ have demonstrated the method of calculation of energy of activation of polymeric substances. The equation used for the calculation of energy of activation (*E*) was,

$$\ln \ln \left(\frac{W_o}{W_t} \right) = \frac{E\theta}{RT_s^2} \quad (1)$$

where $\theta = T - T_s$; W_o is the initial weight; W_t is the weight at any time t ; T_s is the peak temperature; and T is the temperature at particular weight loss. A plot of

$\ln \ln (W_o)/(W_t)$ vs. θ gives an excellent approximation to a straight line. The slope is related to the activation energy. i.e.,

$$\text{Slope} = \frac{E}{RT_s^2} \dots \quad (2)$$

Representative plots are shown in Figure 3. The calculated values for the activation energy of decomposition are listed in Table II. The activation energy associated with each stage of decomposition was also evaluated by the well-known Broido method.²⁵ The

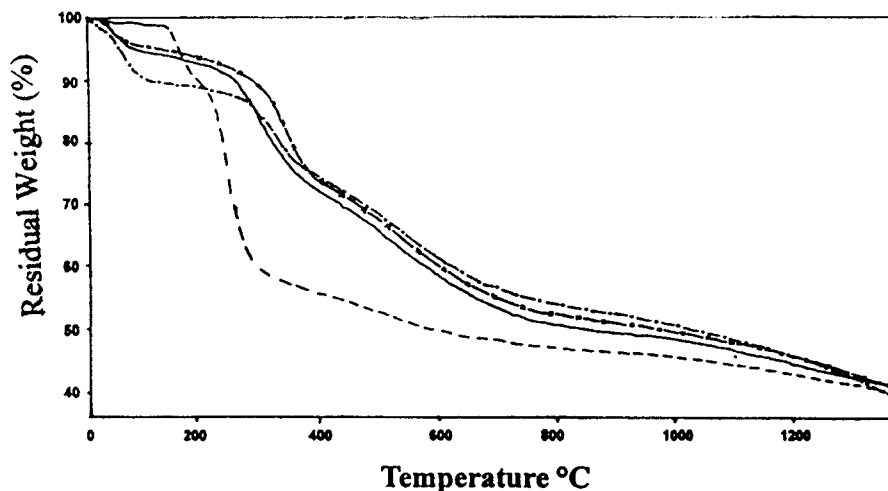


Figure 2 Representative TGA plots of (—) DAPF-ea, (---) DAPF-ap, (-·-) DAPF-en and (-x-) DAPF-pn.

equation used for the calculation of activation energy (E) was,

$$\ln \ln \left(\frac{1}{Y''} \right) = \left(\frac{-E}{R} \right) \frac{1}{T} + \text{Constant} \dots \quad (3)$$

$$\text{where } Y'' = \frac{W_t - W_\infty}{W_0 - W_\infty}$$

that is, Y'' is the fraction of the number of initial molecules not yet decomposed; W_t the weight at any time t ; W_∞ the weight at infinite time (= zero); W_0 the initial weight. A plot of $\ln \ln (1/Y'')$ vs. $1/T$ gives an excellent approximation to a straight line over a range of $0.999 > Y'' > 0.001$. The slope is related to the activation energy. Representative plots are shown in Figure 4. The calculated values for the activation energy of decomposition are listed in Table II.

Differential thermal analyses

The DTA plots of the polymers are given in Figure 5. The Kissinger²⁶ method was used for the determina-

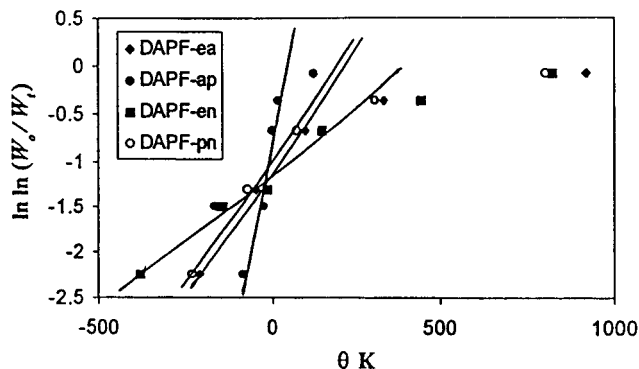


Figure 3 Representative Horowitz and Metzger plots.

tion of activation energy. The equation used for the calculation of activation energy (E) was,

$$\frac{d \left(\ln \frac{\phi}{T_m^2} \right)}{d \left(\frac{1}{T} \right)} = - \frac{E}{R} \dots \quad (4)$$

where ϕ is a constant rate of temperature rise; T_m is the temperature at which the peak differential thermal analyses deflection occurs; T is the temperature. The plot of

$$\left(\ln \frac{\phi}{T_m^2} \right)$$

against the reciprocal of the absolute temperature ($1/T$) gives the energy of activation (E) of the decomposition using the equation,

$$\text{Slope} = - \frac{E}{R} \dots \quad (5)$$

Representative plots are shown in Figure 6. The calculated values for the activation energy of decomposition are 13.86, 9.98, 14.55, and 9.98 kJmol^{-1} for DAPF-ea, DAPF-ap, DAPF-en, and DAPF-pn, respectively. It has been found that all the polymer resins are stable and show maximum peak differential thermal analyses deflection in the temperature range of 350–580°C.

APPLICATIONS

Effect of pH on uptake of metal ion

The effect of pH on metal ion uptake was studied at different pHs by following the recommended proce-

TABLE II
Activation Energy of Decomposition for Various Phenolformaldehyde Supported Chelating Resins

Resin	Peak temperature T_s (°C)	Decomposition temperature range (°C)	% Weight loss	Method	Activation energy (kJmol^{-1})
DAPF-ea	275	100–400	28	HM	15.73
				BR	29.13
	650	400–750	20	HM	15.73
				BR	29.13
DAPF-ap	1000	800–1200	6	HM	6.00
				BR	10.40
	200	100–415	62	HM	31.00
				BR	72.82
DAPF-en	884	415–1200	17	HM	5.30
				BR	7.00
	180	75–240	14	HM	7.11
				BR	18.21
DAPF-pn	475	240–667	31	HM	10.84
				BR	29.13
	973	773–1200	17	HM	6.45
				BR	15.61
DAPF-pn	300	105–400	26	HM	16.54
				BR	16.18
	550	400–700	21	HM	14.08
				BR	31.21
DAPF-pn	950	700–1200	9	HM	3.20
				BR	16.18

HM Horwitz & Metger
BR Broido

Figure 7 indicates that the uptake of metal ion by resins depends on pH. The overall uptake of metal ions by the resin from pH 2 to pH 10 is in the order $\text{Pb(II)} > \text{Hg(II)} > \text{Cr(IV)} > \text{Cd(II)}$. The uptake of Pb(II) and Hg(II) increases with an increase in pH and reaches a plateau value around pH 10. In the case of Cd(II) and Cr(VI) , the uptake increases with pH and reaches a plateau value around pH 6 and decreases with an increase in pH. As uptake of metal ions [Pb(II) and Hg(II)] from basic medium is either high or appreciable, a buffer solution of pH 10 is selected for further studies. In the case of Cd(II) and Cr(VI) , the uptake was maximum at pH 6. Therefore, a buffer solution of pH 6 is recommended for further studies.

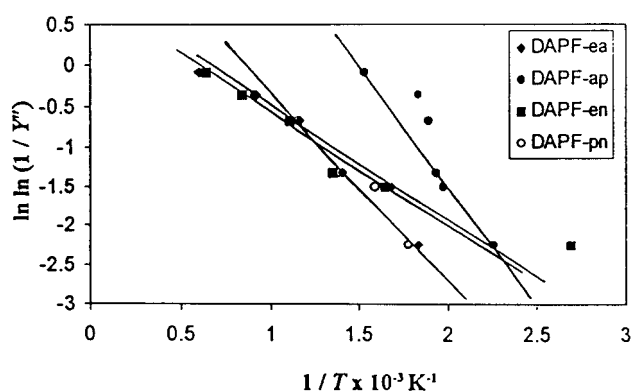


Figure 4 Representative Broido plots.

Effect of time on metal chelation

The effect of time on metal ion uptake was studied at recommended pH by following the procedure described before. From Figure 8 it is evident that 1-h time is sufficient for maximum and constant uptake of the metal ion.

Selectivity studies

The selectivity studies were carried out by following the recommended procedure. In this case, 100 ppm of each metal ion in the solution were allowed to react with the polymer at pH 10. The overall competitive conditions are in good agreement with the results obtained in the noncompetitive conditions. As shown in Figure 9, the metal ion uptake increases from DAPF-ea to DAPF-pn. The high uptake capacity of metal ion by DAPF-pn is probably due to the formation of stable metal chelates. Thus, DAPF-pn is more selective for the removal of Pb(II) , Hg(II) , Cd(II) , and Cr(VI) than the other resins. Thus, the order of selectivity is $\text{DAPF-pn} > \text{DAPF-en} > \text{DAPF-ap} > \text{DAPF-ea}$. The affinity of given polymer for metal ion is in the order $\text{Pb(II)} > \text{Hg(II)} > \text{Cr(VI)} > \text{Cd(II)}$.

Effect of initial concentration of metal ions

An important variable in the uptake of heavy metal ions is its (metal ion) concentration. Figure 10 shows

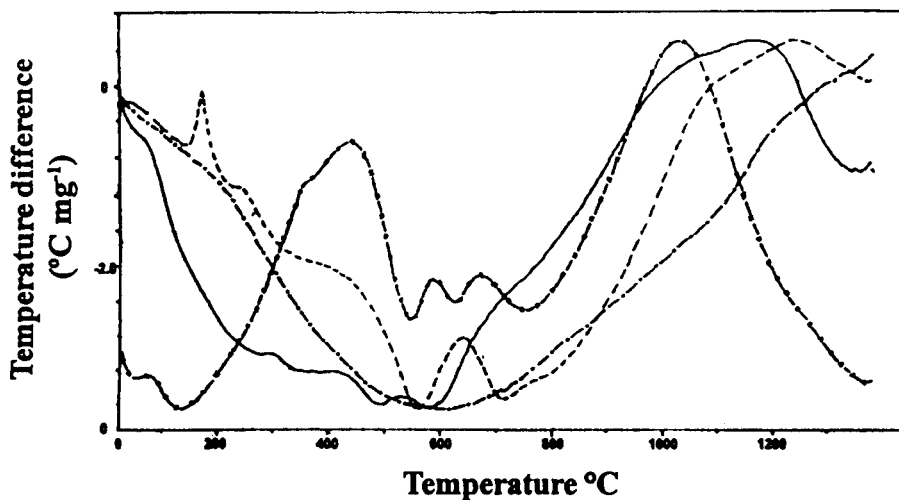


Figure 5 Representative DTA plots of (—) DAPF-ea, (---) DAPF-ap, (-•-) DAPF-en and (-x-) DAPF-pn.

the uptake capacities of the DAPF-ea, DAPF-ap, DAPF-en, and DAPF-pn, for the ions from the single metal aqueous solutions under study. The amount of metal ion uptake was increased with an increase in initial metal ion concentration, and reached a plateau value at higher concentration. The maximum uptake

capacities for DAPF-ea, DAPF-ap, DAPF-en, and DAPF-pn are 10.57, 10.47, 12.45, and 12.87 mg/g of Pb(II); 8.04, 8.42, 8.72, and 11.67 mg/g of Hg(II); 2.24, 2.43, 3.01, and 2.69 mg/g of Cd(II); 3.75, 3.70, 3.69, and 3.53 mg/g of Cr(VI), respectively. The affinity order of

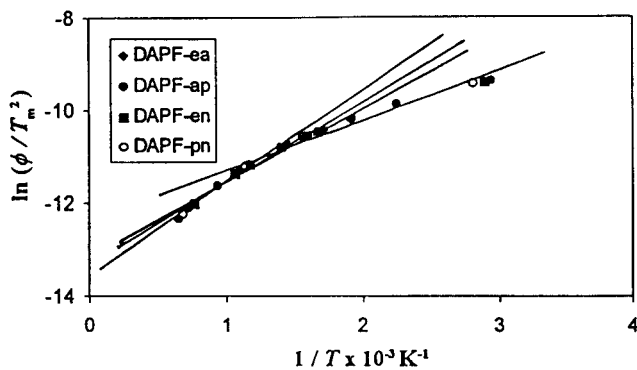


Figure 6 Representative Kissinger plots.

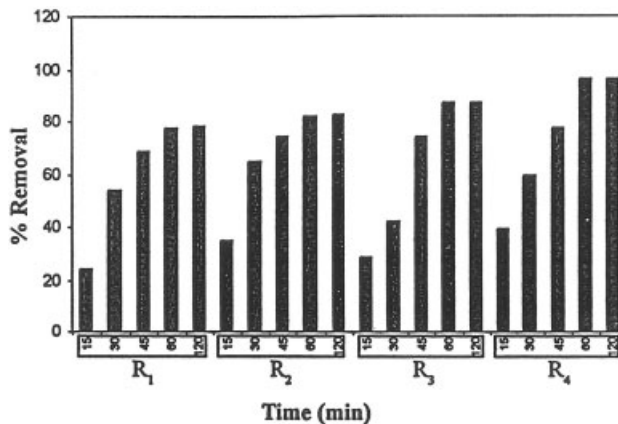


Figure 8 Effect of time on the uptake of Hg(II) by R₁ [DAPF-ea]; R₂ [DAPF-ap]; R₃ [DAPF-en]; R₄ [DAPF-pn].

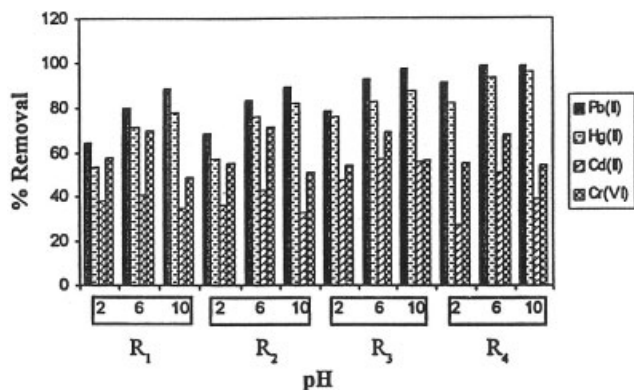


Figure 7 Effect of pH on the uptake of metal ions by R₁ [DAPF-ea]; R₂ [DAPF-ap]; R₃ [DAPF-en]; R₄ [DAPF-pn].

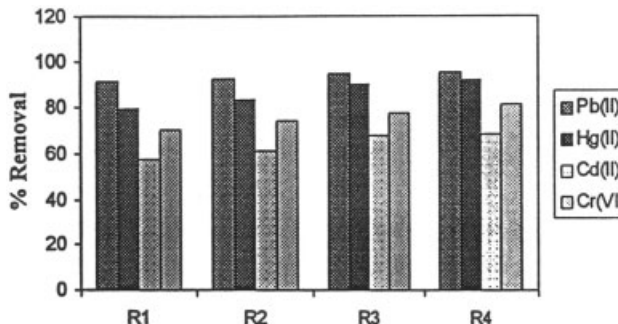


Figure 9 Selectivity on the uptake of metal ions by R₁ [DAPF-ea]; R₂ [DAPF-ap]; R₃ [DAPF-en]; R₄ [DAPF-pn].

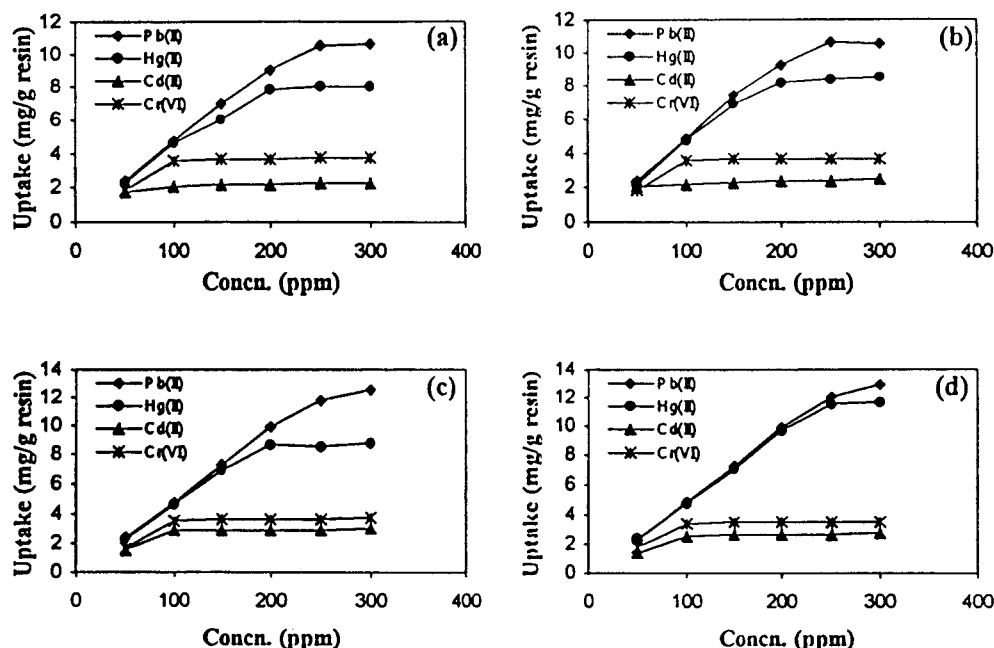


Figure 10 Effect of initial concentration on the uptake of metal ions by (a) [DAPF-ea]; (b) [DAPF-ap]; (c) [DAPF-en]; (d) [DAPF-pn].

the resins is $\text{Pb(II)} > \text{Hg(II)} > \text{Cr(VI)} > \text{Cd(II)}$. Different functionalised polymers with wide range uptake capacities for heavy metal ions have been reported. Kimiaki et al.²⁷ removed 10.3 mg Pb(II) /g resin using a chloromethylated polystyrene containing a $-\text{P(OEt)}_2$ group. Denzili et al.²⁸ showed the uptake of 4.2 mg Cd(II) /g with procion Red MX-3B-immobilized poly (hydroxy-ethylmethacrylate ethylene glycol dimethacrylate). Zhengyu et al.²⁹ found that the Hg(II) equilibrium uptake capacity of HCHO -thiourea-urea copolymer, HCHO - PhOH -urea copolymer and benzoic acid- modified HCHO - thiourea copolymer was 31.5, 30.2, and 28.8 mg/g of resin, respectively. Chattopadhyay et al.³⁰ used a chelating resin containing an imidazolylazo group for the removal of Pb(II) , Ag(II) , and Hg(II) . Thus, metal uptake capacities of present polymers are quite comparable to those employed earlier.

The uptake of metal ions by the resin increases with an increase in initial concentration, and reaches a sat-

uration point at higher concentration. This behavior of the polymers was approximately described by Freundlich's equation

$$Y = KC^{1/n}$$

$$\log_{10} Y = \log_{10} K + 1/n \log_{10} C$$

where Y is the grams of metal ion adsorbed per gram of polymer; C is the initial concentration (mol per liter); n and K are empirical constants.

The plot of $\log_{10} Y$ vs. $\log_{10} C$ gave a straight line. The slope is equal to $1/n$ and the intercept equal to $\log_{10} K$. The K and n values for the adsorption of metal ions on various resins are presented in Table III. Thus, the table suggest that the higher the $1/n$ and K values, the maximum will be the adsorption, and vice versa.

Effect of recyclability on adsorption of metal ions

The most important advantage of the functionalized polymer is its reuse after a particular process. The

TABLE III
The Values of Empirical Constants for the Adsorption of Heavy Metal Ions on Various Chelating Resins Obtained from Freundlich's Equation

Resin	1/n				K			
	Pb(II)	Hg(II)	Cd(II)	Cr(VI)	Pb(II)	Hg(II)	Cd(II)	Cr(VI)
DAPF-ea	0.857	0.750	0.182	0.167	3.162	1.259	0.008	0.010
DAPF-ap	0.938	0.900	0.111	0.091	6.310	3.981	0.005	0.007
DAPF-en	1.000	1.000	0.158	0.094	12.589	10.00	0.004	0.006
DAPF-pn	0.900	0.833	0.090	0.091	3.981	2.818	0.005	0.005

TABLE IV
Recyclability of the Resins on the Uptake of Metal Ions at Recommended pH from 207.2 ppm of Pb(II), 200.6 ppm of Hg(II), 112.4 ppm of Cd(II), and 52.0 ppm of Cr(VI) Solutions

Resin	Cycle No.	Pb(II)		Hg(II)		Cd(II)		Cr(VI)	
		Metal uptake (ppm)	% Desorption	Metal uptake (ppm)	% Desorption	Metal uptake (ppm)	% Desorption	Metal uptake (ppm)	% Desorption
DAPF-ea	1	182.7	94.9	156.3	98.3	45.6	96.6	45.9	96.2
	2	180.5	96.1	156.2	97.6	45.2	97.2	44.8	97.8
	3	179.4	98.8	154.3	98.2	44.5	96.9	44.2	96.3
	4	178.3	98.2	155.2	96.5	44.1	98.4	43.5	98.4
DAPF-ap	1	185.1	96.7	164.6	95.9	48.5	98.6	45.4	98.2
	2	184.4	97.3	162.3	97.3	46.5	97.8	44.8	97.9
	3	184.1	98.6	161.4	96.8	45.8	97.3	44.2	97.8
	4	183.3	96.4	160.2	98.9	45.3	96.1	43.6	96.3
DAPF-en	1	200.8	94.5	174.6	97.7	64.0	98.7	45.9	98.5
	2	200.2	98.5	174.0	96.9	63.5	97.5	45.2	97.1
	3	198.4	99.4	173.4	98.2	63.1	98.6	43.5	98.7
	4	196.3	96.3	173.2	95.5	62.8	98.4	43.1	98.4
DAPF-pn	1	204.7	97.6	192.6	97.4	57.4	97.3	45.2	97.6
	2	203.4	96.5	190.9	97.3	56.6	96.2	44.4	96.8
	3	202.2	96.8	190.2	96.6	58.1	96.6	44.1	96.2
	4	201.6	98.2	189.4	98.2	55.5	98.4	43.7	98.5

polymer used can be brought into its original state by desorbing the metal ions from metallated resins using 6 M HCl in tetrahydrofuran. The metal free polymers can be reused after neutralization. In acidic medium the metallated polymers are protonated and releases the metal ion into solution. For every cycle the metal ion uptake and % desorption was studied. The data are given in Table IV. The metal ion uptake was found to be almost the same even after four cycles using different polymers.

Industrial applications

The resins have been successfully employed for the removal of heavy metal ions from industrial waste water. The industrial waste water was collected from Karnataka pollution control board, Bangalore. The data are given in Table V.

TABLE V
Removal of Heavy Metal Ions from Industrial Waste Water^a

Resin	Heavy metal ion removed ^b			
	Hg(II)	Cd(II)	Cr(VI)	Ni(II)
DAPF-ea	0.290	0.625	1.135	0.410
DAPF-ap	0.310	0.705	1.170	0.180
DAPF-en	0.355	0.810	1.225	0.640
DAPF-pn	0.420	0.780	1.310	0.680

^a Certified composition of the sample (in ppm) 12.33 Hg; 22.8 Cd; 30.9 Cr; 18.6 Ni.

^b mg of metal per gram resin.

CONCLUSION

Functionalized polymers have been synthesized by employing a condensation polymerization technique and subsequent modifications. These polymers have been characterized, based on elemental analyses, infrared, and ¹H-NMR spectra. Thermal and kinetic data suggest that the resins are highly stable, and may be used for the removal of metal ions at room temperature. The metal uptake efficiency increases with pH and reaches a plateau value around pH 10 [for Pb(II) and Hg(II)]. However, the uptake efficiency reaches a plateau value around pH 6 (for Cd(II) and Cr(VI)) and decreases with an increase in pH. The favorable characteristic of present polymers is the time required for the maximum and constant uptake of metal ion from aqueous media. Just 1-h time is sufficient for the removal of metal ion in the present methods. As observed in Figure 10, the uptake was increased with concentration and reached a plateau value. This is because at higher concentrations no more free sites will be available on chelating polymeric surface either for adsorption or for chelation. The metal ion selectivity of present functionalized polymers is found to be in the order Pb(II) > Hg(II) > Cr(VI) > Cd(II). The metal ion uptake efficiency of the polymers is not altered much even after five cycles. The metal ion uptake efficiency of the polymers are in the order, DAPF-pn > DAPF-en > DAPF-ap > DAPF-ea. Thus, the polymer having propylenediamine moiety (DAPF-pn) is found to be more efficient in the removal of heavy and toxic metal ions.

The authors thank SIF, Bangalore, and NMDC, Hyderabad, for providing spectral and AAS data, respectively.

References

1. Huang, C. P.; Bstarz, F. B. *J Environ Eng* 1971, 104, 863.
2. Wing, R. E.; Doone, W. H.; Russel, C. R. *J Appl Polym Sci* 1975, 19, 847.
3. Netzer, A.; Wilkinson, P.; Beszedits, S. *Water Res* 1974, 8, 813.
4. Sundersen, B. B.; Bulusis, K. R.; Kulkarni, D. M. *Indian J Environ Health* 1978, 20, 413.
5. Patterson, J. W.; Allen, H. F.; Scala, J. J. *J Water Pollut Control Fed* 1977, 9, 2397.
6. Lindstdt, K. K.; Hanch, C. P.; Connor, J. J. *J Water Pollut Control Fed* 1971, 43, 1507.
7. Marinsky, J. A.; Anspach, W. M. *J Phys Chem* 1975, 79, 439.
8. Rivas, B. L.; Maturana, H. A.; Villegas, S. *Polym Bull* 1997, 39, 445.
9. Rivas, B. L.; Pooley, S. A.; Maturana, H. A.; Villegas, S. *J Appl Polym Sci* 2001, 80, 2123.
10. Dominguez, L.; Benak, K. R. *Economy, J. Polym Adv Technol* 2001, 12, 197.
11. Abrams, I. M.; Miller, J. R. *React Polym* 1997, 35, 7.
12. Warshawsky, A.; Marques, G.; Kahana, N.; Kampel, V.; Bourdelande, J. L.; Valiente, M. *Macromol Chem Phys* 2001, 202, 2659.
13. Trochimczuk, A. W.; Kolarz, B. *Eur Polym J* 2000, 36, 2359.
14. Saraydin, D.; Isikver, Y.; Sahiner, N. *Polym Bull* 2001, 47, 81.
15. Loh, X. J.; Deen, R.; Gan, Y. Y.; Gan, L. H. *J Appl Polym Sci* 2001, 80, 268.
16. Reddy, K. H.; Reddy, A. R. *Indian J Chem* 2002, 41A, 934.
17. Reddy, A. R.; Reddy, K. H. *Indian J Chem Technol* 2002, 9, 97.
18. Reddy, K. H.; Reddy, A. R. *J Appl Polym Sci* 2003, 88, 414.
19. Reddy, A. R.; Reddy, K. H. *Proc Indian Acad Sci* 2003, 115, 155.
20. Reddy, K. H.; Lingappa, Y. *Indian J Chem* 1998, 36A, 1130.
21. Reddy, K. H.; Lingappa, Y. *Ind J Chem* 1994, 33A, 919.
22. Cooper, S. R.; *Org Syn* 1941, 21, 103.
23. Marczenko, Z. *Spectrophotometric determination of Elements*; Wiley: New York, 1976, p. 355.
24. Horowitz, H. H.; Metzger, G.; Esso Res Engg Co., Linden, NJ, *J Anal Chem* 1963, 35, 1464.
25. Broido, A. *J Polym Sci* 1969, 7, 1761.
26. Kissinger, H. E. *Natl Bur Standards J Anal Chem* 1957, 29, 1703.
27. Kimiaki, M.; Masahiro, A. *Chem Abstr* 1998, 109, 134491.
28. Denzili, A.; Buyuktuncel, V.; Genc.; Piskin, E. *Anal Lett* 1998, 31, 2791.
29. Zhengyu, Z. *Huaxueshijie* 1987, 28, 252.
30. Chattopadhyay, P.; Sinha, C.; Pal, D. K. *J Anal Chem* 1997, 357, 368.